



# Article Historical Zooplankton Composition Indicates Eutrophication Stages in a Neotropical Aquatic System: The Case of Lake Amatitlán, Central America

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Abstract: This paper presents a study of freshwater zooplankton biodiversity, deemed as a reliable indicator of water quality. The Guatemalan Lake Amatitlán, currently used as a water source, has shown signs of progressive eutrophication, with perceptible variations of the local zooplankton diversity. Biotic and abiotic parameters were determined at four sites of Lake Amatitlán (Este Centro, Oeste Centro, Bahía Playa de Oro, and Michatoya) in 2016 and 2017. The local composition, the species richness and abundance of zooplankton, and the system environmental parameters were analyzed during both years surveyed. Biological data suggesting eutrophication of this tropical system were obtained, including a high rotifer abundance (11 species: the rotifers *Brachionus havanaensis* (109 ind  $L^{-1}$ ) and *Keratella americana* (304 ind  $L^{-1}$ ) were the most abundant species in this lake). The presumably endemic diaptomid copepod species, *Mastigodiaptomus amatitlanensis*, was absent in our samples, but we report the unprecedented occurrence of two Asian cyclopoid copepods (i.e., *Thermocyclops crassus* and *Mesocyclops thermocyclopoides*) for Lake Amatitlán and Guatemala. The presence of larger zooplankters like adults and immature copepods (i.e., *Arctodiaptomus dorsalis*) and cladocerans (*Ceriodaphnia* sp.) at site "Este Centro" indicates a relatively healthy zooplankton community and represents a focal point for managing the conservation of this lake.

Keywords: conservation; eutrophication; exotic species; tropical lakes; zooplankton

# 1. Introduction

The knowledge of zooplankton in the Neotropical region is growing with fragmented studies. Therefore, it is likely that the species richness of zooplanktonic taxa is underestimated because of the presumably high diversity and scarcity of zooplankton taxonomists [1–3]. In addition, the progressive destruction of aquatic habitat and the progressive spread of exotic species threaten native biodiversity, ecosystem health, and environmental services.

The zooplankton community and abundance are closely linked to the trophic state of the water system; for this reason, its diversity has been deemed as an indicator of water quality [4]. In eutrophicated systems (at tropical and temperate latitudes), the dominance of microzooplankton is common, compared with larger organisms, owing to the increased availability of food and water conditions [5,6].



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**Copyright:** © 2021 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). For four decades, the Guatemalan Lake Amatitlán has shown signs of progressive eutrophication related to anthropic factors (i.e., peripheral population growth and urbanization, intensive use of water for agricultural irrigation), thus promoting the advancement towards eutrophication, related to the input of nearly 50% of the untreated residual urban and industrial waters from Guatemala City [7–10]. Because of this, some actions have been proposed to address this problem, either from the governmental level (i.e., Autoridad para el Manejo Sustentable de la cuenca del lago Amatitlán, AMSA 1996) or from descriptive studies of the lake involving the lake zooplankton biodiversity, like those by Basterrechea-Díaz (1997) [7] and Brandorff (2012) [11]; however, studies related with tropical epicontinental waterbodies have been more focused on environmental factors rather than biological community attributes or general limnology [12,13]; thus, the zooplankton biodiversity in Guatemala remains largely unknown [14], with only a few studies in Guatemalan lakes [15,16]. Most studies in Lake Amatitlán and Guatemala are more focused on current data instead of historical analysis.

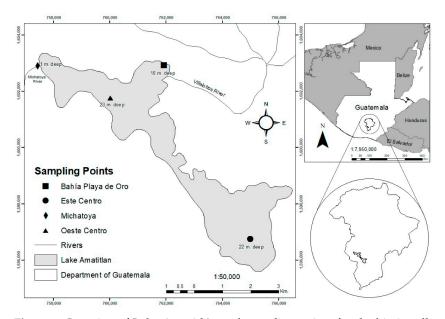
Based on the analysis of both, historical and current data of zooplankton biodiversity and environmental conditions of Lake Amatitlán, we present information on the zooplankton distribution, species richness, abundance, and its relation with successive changes of its trophic state.

# 2. Materials and Methods

# 2.1. Study Sites and Sampling Methods

Lake Amatitlán is the fourth largest lake in Guatemala, Central America, and one of the most emblematic waterbodies of this country. This lake is a warm monomictic waterbody in the highland of Guatemala, located at an altitude of 1186 m above sea level (m.a.s.l.), with an area of 15.2 km<sup>2</sup> and 11 km length and a maximum depth of 23 m. Its formation originated from volcanic activity of Pacaya, Fuego, and Agua in the late Quaternary [10,14,17].

Four sampled sites were considered: Este Centro (EC), Oeste Centro (OC), Bahía Playa de Oro (BPO), and Michatoya (MICH) to analyze the zooplankton species that inhabit the eastern and western regions of Lake Amatitlán (Figure 1). The latter two sites (BPO and MICH) are in the runoff of Villalobos and Michatoya rivers, respectively [14]. Water samples for biotic and abiotic variables were collected for 2016 and 2017 in the rainy (May–October) and dry seasons (November–April).



**Figure 1.** Location of Lake Amatitlán and sampling points for the biotic collection methodology. Water-filtered, vertical, and horizontal trawls as defined by Cervantes-Martínez & Gutiérrez-Aguirre (2015) [2].

#### 2.1.1. Species Richness

Zooplankton samples (n = 8) were collected by vertical and horizontal trawls with a 45 µm plankton net between 1 and 22 m depth to ensure representative samples to evaluate the species richness in the lake, as it is well known that zooplankton tends to have vertical and horizontal migrations [2].

## 2.1.2. Species Abundance and Abiotic Variables

To estimate the zooplankton abundance, a known volume of water between 30 and 100 L was filtered through a 45  $\mu$ m zooplankton net. The water was determined with a 2.1 L<sup>-1</sup> capacity Van Dorn bottle [2,18]. Species abundance was determined by the account of two main groups: Rotifera and Copepoda, present in three aliquots of 1 mL each from the filtered samples, then the data were standardized as individuals per liter (ind L<sup>-1</sup>) in each sampled site [19].

Abiotic variables were measured in situ monthly for both years of study and in all the water columns, with the multiparametric proves WTW Cond 197i, WTW Oxi 1970i, and HACH HQ for water temperature (°C), pH, oxygen concentration  $O_2$  (mg L<sup>-1</sup>), total dissolved solids (mg L<sup>-1</sup>), and conductivity ( $\mu$ S cm<sup>-1</sup>). With the actual environmental, richness, and zooplankton abundance data, a description of the trophic state of Lake Amatitlán was proposed.

# 2.2. Historical and Actual Records of Zooplankton and Environmental Parameters Analysis

Specific classification of Rotifera, Cladocera, and Copepoda of recently collected samples (collected in 2016 and 2017) was done according to Koste (1978) [20], Fontaneto & De Smet (2015) [21], Elías-Gutiérrez et al., (2008) [22], and Suárez-Morales et al., (2020) [23].

The presence/absence of the current zooplankton inventory was compared with previous surveys by Juday (1915) [24], Basterrechea-Díaz (1997) [7], and the record of copepods from the previous surveys of Wilson (1941) [25] and Brandorff (2012) [11], in order to analyze the historical composition of zooplankton of Amatitlán lake.

Historical environmental data recorded by Juday (1915) [24], Brezonik & Fox (1974) [26], Basterrechea-Díaz (1997) [7], and Ellenberg (2014) [27] were compared with the current data surveyed in this study.

## 3. Results

#### 3.1. Species Richness

A total of 15 species of zooplankters including rotifers and crustaceans were found in the lake for 2016–2017 (Table 1); rotifers showed the highest species richness (80% of zooplankton species recorded), while copepods represented 20% of all zooplankton species in the lake.

We provide the first record of two cyclopoid exotic species (*Mesocyclops thermocyclopoides* and *Thermocyclops crassus*) for Lake Amatitlán and Guatemala. The endemic calanoid copepod, *Mastigodiaptomus amatitlanensis*, was absent in our current survey and the record of *Arctodiaptomus dorsalis* in Lake Amatitlán was confirmed here. Cladoceran crustaceans were very scarce in our samples; only a single specimen of *Ceriodaphnia* sp. was observed. The Brachionidae was the family with the highest species richness among rotifers in 2016 and 2017 (Table 1).

Nowadays, the east region (site EC) of Lake Amatitlán had the highest species richness in the lake (14 species), compared with the western region (9 species including the exotic *T. crassus* at OC). The largest zooplankters of the lake, including the cladoceran *Ceriodaphnia*, (~2 mm) [21], the calanoid copepod *A. dorsalis*, and the cyclopoid copepod *M. thermocyclopoides*, occurred in eastern region.

**Table 1.** Current and historical records of zooplankton species richness in Lake Amatitlán. Currently recorded species are shown in columns (1) EC, (2) OC, (3) BPO, and (4) MICH. Historical records are shown in columns 5–8, following data by Brandorff (2012) [11]; Basterrechea-Díaz (1997) [7]; Wilson (1941) [25]; and Juday (1915) [24], respectively. Presence (x), absence (-), new records (\*).

Species –	Current Data				Historical Data			
	1	2	3	4	5	6	7	8
Phylum: Rotifera Monogononta: Ploimida								
Family: Epiphanidae Harring, 1913								
<i>Epiphanes macroura</i> Barrois & Daday, 1894 *	х	х	-	-	-	-	-	-
Family: Brachionidae Ehrenberg, 1838	-	-	-	-	-	-	-	-
Anuraeopsis fissa (Gosse, 1851) *	х	-	-	-	-	-	-	-
Brachionus angularis (Gosse, 1851) *	x	х	х	х	-	-	-	-
B. calyciflorus Pallas, 1766 *	x	x	x	x	_	-	-	-
<i>B. plicatilis</i> Müeller, 1786 *	x	-	-	x	_	-	-	-
B. havanaensis Rousselet, 1911 *	x	х	х	x	-	-	_	-
Keratella sp.	-	-	-	-	_	х	-	-
K. americana Carlin, 1943 *	х	х	х	х	-	-	_	-
K. cochleraris (Gosse, 1851)	-	-	-	-	-	_	_	x
Family: Trichocercidae Harring, 1913								X
<i>Trichocerca</i> cf. <i>longiseta</i> (Schrank, 1802) *	_	Y	х	v				
<i>T. pusilla</i> (Lauterborn, 1898) *	-	x		x	-	-	-	-
Family: Asplanchnidae Eckstein, 1883	-	х	х	х	-	-	-	-
Asplanchna sieboldi (Leydig, 1854) *	Ň	Y	Ň					
Flosculariaceae: Family: Trochosphaeridae	х	х	х	-	-	-	-	-
Harring, 1913								
Filinia longiseta (Ehrenberg, 1834)	x	x	X	х	-	-	-	Х
<i>F. terminalis</i> (Plate, 1886) *	х	х	х	-	-	-	-	-
Subclass: Bdelloidea *	х	-	-	-	-	-	-	-
Superclass: Crustacea Brachiopoda:								
Cladocera: Anomopoda								
Family: Daphniidae Straus, 1820								
Daphnia sp.	-	-	-	-	-	х	-	-
D. hyalina Leydig, 1860	-	-	-	-	-	-	-	X
Ceriodaphnia sp.	х	-	-	-	-	х	-	x
C. lacustris Birge, 1893	-	-	-	-	-	-	-	Х
C. pulchella Sars, 1862	-	-	-	-	-	-	-	x
Family: Bosminidae Sars, 1865								
<i>Bosmina</i> sp.	-	-	-	-	-	x	-	-
Bosmina longirostris O. F. Müeller, 1776	-	-	-	-	-	-	-	x
Family: Chydoridae Stebbing, 1902								
Chydorus sphaericus (O.F. Müeller, 1785)	-	-	-	-	-	-	-	x
Copepoda: Calanoida Family:								
Diaptomidae G.O. Sars, 1932								
Subfamily: Diaptominae Kiefer, 1932								
Arctodiaptomus dorsalis (Marsh, 1907)	х	-	-	-	х	-	-	-
Mastigodiaptomus albuquerquensis (Herrick,								
1895)	-	-	-	-	-	-	-	Х
M. amatitlanensis (Wilson, 1941)	-	-	-	-	-	-	х	-
Copepoda: Cyclopoida Family: Cyclopidae								
Kiefer, 1927								
Subfamily: Cyclopinae Kiefer, 1927								
<i>Thermocyclops crassus</i> (Fischer, 1853) *	-	х	-	-	-	-	-	-
Mesocyclops thermocyclopoides Harada, 1931 *	х	-	-	-	_	-	-	-
Subfamily: Eucyclopinae Kiefer, 1927								
Eucyclops serrulatus (Fischer, 1851)	-	-	_	-	-	-	-	x
Nauplii	x	x	x	x	-	x	_	x
Juvenile Cyclopoid	x	x	x	x	-	-	_	~
Juvenile Cyclopold					-	-	-	-
juverine Calanoiu	Х	x	Х	Х	-	-	-	-

Our revision of the zooplankton community (Table 1) indicates that the historical data presented a great microcrustacean richness with the record of eight cladoceran species (*Dapnia* sp., *D. hyalina*, *Ceriodaphnia* sp. *C. lacustris*, *C. pulchella*, *Bosmina* sp., *B. longirostris*, and *Chydorus sphaericus*) and the three calanoid copepods: *A. dorsalis*, *Mastigodiaptomus albuquerquensis*, and the endemic *M. amatitlanensis*. The historical record of rotifers had the lowest species richness including three monogonont species. In our survey, the rotifer species richness increased significantly with 12 species not hitherto reported from the lake, including the record of organisms from the Subclass Bdelloidea.

#### 3.2. Species Abundance

In this study, the total rotifer abundance was 522.7 ind  $L^{-1}$ . Rotifers represent the most abundant group in the lake; their numerical abundance is considerably higher than that recorded for copepods, including immature stages (7.1 ind  $L^{-1}$ ). Cladocerans were almost absent from our samples.

Species with the highest abundance at all sites were the rotifers *B. havanaensis* (109 ind  $L^{-1}$ ) and *K. americana* (304 ind  $L^{-1}$ ), with a considerably lower abundance in the eastern area (9.3 and 121.8 ind  $L^{-1}$ , respectively). Species of the family Brachionidae were the most abundant mainly in the western region (sites OC, BPO, and MICH), whereas the lowest abundance of rotifers occurred in the eastern region (site EC) (see Table 2).

<u>Caracian</u>	Abundance (ind L <sup>-1</sup> )					
Species	EC	OC	BPO	MICH		
Brachionusangularis	0.00	0.00	71.56	1.87		
Brachionuscalyciflorus	0.70	0.70	85.56	4.67		
Brachionusplicatilis	0.00	0.47	0.00	0.00		
Trichocerca cf. longiseta	0.00	0.00	14.00	12.60		
Trichocercapusilla	0.00	8.40	13.22	11.20		
Asplanchnasieboldi	1.40	1.17	2.33	0.93		
Filinialongiseta	0.00	1.40	28.00	13.07		
Filiniaterminalis	8.17	49.23	35.00	60.20		
Brachionushavanaensis	9.33	153.53	108.89	165.20		
Keratellaamericana	121.80	432.60	265.22	408.33		
Nauplii	3.50	2.57	3.11	1.40		
Juvenile Cyclopoid	5.83	1.63	1.56	0.93		
Juvenile Calanoid	2.80	0.47	3.89	0.47		
M. thermocyclopoides	0.23	0.00	0.00	0.00		

**Table 2.** Abundance (ind  $L^{-1}$ ) calculated from zooplankton samples for all the studied points of Lake Amatitlán in 2017.

The local copepod abundance was represented mainly by nauplii and juvenile stages of Calanoida and Cyclopoida (average = 2.6, 2.5 and 1.9 ind  $L^{-1}$ , respectively), values resembling those recorded for the Rotifera like *B. plicatilis* (1.1 ind  $L^{-1}$ ) and *A. sieboldi* (2.3 ind  $L^{-1}$ ) in all the study sites, compared with adult copepods, where the abundance of the adult *M. thermocyclopoides* present only in EC was 0.23 ind  $L^{-1}$ .

#### 3.3. Environmental Variables

Environmental variables values in both analyzed years, in general, presented basic pH values (>8  $\pm$  0.33), dissolved oxygen showed an average of 4.76  $\pm$  5.21 and 4.65  $\pm$  4.92 mg L<sup>-1</sup>, whereas temperature averaged 24  $\pm$  1.31 °C, conductivity presented average values of 655.95  $\pm$  59.52 and 678.23  $\pm$  68.29  $\mu S$  cm<sup>-1</sup>, and finally TDS showed average values of 339.99  $\pm$  47.35 and 341.43  $\pm$  30.30 mg L<sup>-1</sup>, respectively (Table 3).

Environmental Variables	1910 [24]	1969 [ <mark>26</mark> ]	1985–1995 [7]	2008–2013 [27]	2016 ***	2017 ***
pH	ND	7.70	7.75	8.69	8.26	8.33
Water temperature (°C)	19.86	ND	22.75	25.00	24.46	24.23
Conductivity ( $\mu$ S cm <sup>-1</sup> )	ND	830	802	682	655.95	678.23
TDS (mg $L^{-1}$ )	ND	ND	610	ND	339.99	341.43
Dissolved oxygen $O_2$ (mg L <sup>-1</sup> )	4.74 *	8.40 **	4.20	8.90	4.76	4.65

**Table 3.** Historical and current environmental mean data of the water column recorded by previous surveys and this study. Juday (1915) [24], Brezonik & Fox (1974) [26], Basterrechea-Díaz (1997) [7], Ellenberg (2014) [27]. ND: no data available.

\* Data originally recorded in cubic centimeters per liter of water. \*\* Original data recorded at surface. \*\*\* Data recorded in this study.

The historical data presented in Table 3 show pH with slightly neutral values in 1969 to 1985–1995, whereas in the first two decades of the XXI century, the pH increased to reach clearly basic values, over 8. The water temperature changed along the time, 19.86 °C in 1910 to 24.23 °C in 2017. Conductivity and total dissolved solids decreased on average by 18.29 and 44.10%, respectively.

#### 4. Discussion

The environmental parameters surveyed in this study can show the progressive eutrophication on Lake Amatitlán, according to the historical data recorded by authors like Juday (1915) [24], Brezonik & Fox (1974) [26] Basterrechea-Díaz (1997) [7], and Ellenberg (2014) [27]. The historical change in environmental and biological variables could reveal strong evidence of the current eutrophication of this lake. For instance, the observed changes of pH values, that is, an average of 8.26 and 8.33 in 2016–2017, differ in contrast from the values recorded in 1969 (7.70) [26], 1985–1995 (7.75) [7], and 2008 (9.3) [17].

The basic pH and the high concentration of dissolved oxygen at the surface promoted an increase of microzooplankters, like rotifers (especially *B. havanaensis* and *K. americana*), and a decrease of larger species like cladocerans and adult copepods, indicators of the system trophic state per se. Similar conditions have been recorded in American eutrophicated subtropical and tropical water bodies [4,28,29] as well as in other water bodies (i.e., temperate coastal water bodies) in which the replacement of larger copepod with smaller ones has been reported to the result from the eutrophication process [6].

Recently, phytoplankton blooming has been described as a consequence of this eutrophication progress in Lake Amatitlán, presenting a high concentration mainly in *Microcystis* sp. and *Dolichospermum* sp. cyanobacteria preceded by the diatom algae *Niszcha* sp. at the surface of the lake [9], which in turn allows herbivorous zooplankters like brachionid rotifers to become dominant organisms in eutrophicated epicontinental waterbodies [20].

In earlier studies on Lake Amatitlán, the zooplankton community was largely dominated by cladocerans and copepods. In 1915 [24], zooplankton had a widely different composition compared with our results: rotifers were then the less abundant zooplankton group in the lake (0.3 ind L<sup>-1</sup>), preceded by copepods (11.6 ind L<sup>-1</sup>) and cladocerans, the most abundant zooplankton group at that time (14.4 ind L<sup>-1</sup>). The system trophic state is also related to the zooplankters body size; that is, a stronger level of eutrophication is frequently expressed by a greater abundance and species richness of microzooplankters like small rotifers [4,6,28,29]. A possible explanation of the local absence or scariness of larger zooplankters (i.e., *Ceriodaphnia* sp., adult cyclopoid and calanoid copepods, including *M. amatitlanensis*) could result from the competition for available food [5], eventually explaining the strong dominance of small brachionid herbivorous rotifers like *B. havanaensis* and *K. americana*.

The presence and high abundance of these latter species, together with another species of *Brachionus* and *Keratella* at the east region of Lake Amatitlán, suggest that eutrophic conditions that make food available for these microphagous species [30].

In the case of *A. dorsalis*, this species is widespread in America [31] and has been recorded as an invasive exotic copepod in Asiatic waterbodies [32,33]. The environmental conditions of Lake Amatitlán seem to be adequate for the development of this species be-

cause it shows a selective feeding on phytoplankton; thus, it frequently inhabits moderately to strongly eutrophicate environments [31,32], like Amatitlán lake.

It is well known that many diaptomid copepods tend to have restricted distributional patterns and endemic distributions in neotropical lakes [34]. Then, the local absence of the endemic copepod *M. amatitlanensis* in this study could be another indicator of the progressive eutrophication of Lake Amatitlán, because, since its description by Wilson (1941) [25], this species has not been recorded in other regional studies (i.e., Elías-Gutiérrez et al., 2008 [35]; Brandorff, 2012 [11]; and Gutiérrez-Aguirre, et al., 2020 [36]). It is probable that *M. amatitlanensis* occurs in other lakes of Guatemala (or Central America) and it is expected to be collected from adjacent systems. It is also probable that this species dwells at higher depths not easily reached by standard nets.

Our results showed a clear zonification; the eastern region (site EC) diverges from the other sites because of the absence of adjacent rivers (see Figure 1), its distance from the other sampling points (the closest site is OC, 7.04 km away), and its separation from other sites owing to a train riel that divides the lake in two [14]. Therefore, the EC area has the best conservation status of the lake, precisely where we found the greatest species richness and the larger zooplankters, with the copepods *T. crassus* (average body length of 0.56–0.93 mm) [37], *M. thermocyclopoides* (0.78–0.89 mm) [38], and *A. dorsalis* (0.77–1.13 mm) [31] among them. Thus, it is convenient to consider EC as a potential conservation site as it has better environmental conditions for the conservation and preservation of zooplankton biodiversity.

On the other hand, we report the presence of two exotic cyclopoid copepod species for the Central American Lake Amatitlán and Guatemala country, *M. thermocyclopoides* and *T. crassus. M. thermocyclopoides* is a native species from Taiwan and is well spread in Asia and Africa, and commonly widespread at tropical latitudes. This species has been recorded in lakes from South Mexico in epicontinental waterbodies from Chiapas state, Mexico, considering that their introduction may be related to anthropic factors (i.e., agriculture and aquaculture) [37,38]. This is the second record of the invasion of this species in Central American countries, as it has been recorded before in Costa Rican water bodies by Collado et al. (1894) [39], and the ecological potential of *Mesocyclops* use as biocontrol of vector mosquitoes like *Aedes aegypti* is well known [40–42]. Therefore, its finding in Guatemalan lakes represents a source for mass culture of this copepod to be used as biocontrol.

*Thermocyclops crassus* is commonly spread at tropical latitudes in Africa, Australia, and Asia; it was also recorded in Laurentian great lakes in the United States of America [43]; recorded for the first time in tropical lakes from Tabasco state, Mexico [37]; as well as in small ponds of San José Province in Costa Rica [39]. Being a thermophilic species, *T. crassus* has a narrow temperature tolerance [44], so it may be a local indicator of the temperature changes in the lake along time.

Finally, the physical, chemical, and biological conditions of the lake have clearly changed over time, from being a lake with oligotrophic characteristics to one with hyper-trophic conditions in a relatively short period of time (100 years, approximately), allowing us to follow and describe the stages and speed of the eutrophication process of a large neotropical lake.

# 5. Conclusions

The historical analysis of zooplankton composition in the lake presented in this study reinforces the knowledge of its eutrophic state, suggesting a useful role of the zooplankton as a bioindicator and making possible the visualization of the changes in its composition over time, showing the progressive trophic state towards eutrophic or hypereutrophic conditions.

It is likely that the absence of the endemic species *M. amatitlanensis* is a warning sign regarding the accelerated loss of biodiversity and reinforces the idea that zooplankton is a great tool as a bioindicator of the health status for continental aquatic ecosystems, in both tropical and temperate latitudes.

Further studies analyzing bottom sediments to search resting eggs of zooplankton in Lake Amatitlán and around it can answer the question of the absence of *M. amatitlanensis*, where this type of knowledge is also scarce in inland aquatic systems of the region.

Finally, is convenient to consider the isolated site EC as a focal point for conservation as it presents better environmental conditions for the conservation and preservation of zooplankton biodiversity, owing to the record of the largest zooplankters found in this site.

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